

16 **Abstract**

17 Timber and non-timber ecosystem services (ESs) of forests can have trade-offs. These trade-offs are
18 often influenced by local characteristics, and a higher awareness of local ESs among the location
19 population could support forest management to supply ESs sustainably. This study examines trade-
20 offs among timber and non-timber ESs in three adjacent municipalities in Japan where social contexts
21 differ and discusses them in relation to the environmental awareness of each community. First, we
22 explored the local awareness of the population of ESs in interviews. Then we produced maps of
23 landslide prevention, sediment retention, and forest recreation ESs in plantations at 30-m resolution
24 and classified forests according to evaluations of each ES. We overlaid the ES maps with a map of
25 logging locations from the previous 5 years to calculate the logged ratio for each ES class. In a region
26 with a long history of forestry, where awareness of ESs seems to be high, forests providing a wider
27 range of ESs had a lower logged ratio than forests with lower ESs. In contrast, in a region in which
28 contracted foresters from outside of that region were coordinating logging activities, even forests
29 providing numerous ES had a high logged ratio. Thus, increasing awareness of ESs amongst the local
30 population may lead to a more balanced use of ESs. Our results indicate that local governments would
31 be best placed to raise awareness by educating forest enterprises or providing science-based
32 information on ESs to foresters. We conclude that analyses of local ES trade-offs under consideration
33 of the social context as presented here, is the first step towards developing and maintaining
34 sustainable forest management principles.

35 **1. Introduction**

36 Since our daily lives benefit from ecosystem services (ESs) both directly and indirectly, the
37 continuous maintenance and improvement of ESs is an important goal (Millenium Ecosystem
38 Assessment, 2005). ESs vary by region, influenced by diverse landscapes shaped by local weather and
39 topography (Frizzle et al., 2022; Katila et al., 2020). Simultaneously, their use by local people also
40 varies (Karjala et al., 2004; Sherry et al., 2005). In some regions, ecosystem-based disaster risk
41 reduction is vital (Dalimunthe, 2018); in others, local people make a living from the commercial use
42 of natural products (Santika et al., 2019). People who rely on local ESs evolve their lifestyles to

43 benefit from them more efficiently (Asah et al., 2014; Orenstein and Groner, 2014). Such local use can
44 affect ES supply (Barlow et al., 2016; Gardner et al., 2010). Thus, the local supply and use of ESs
45 have woven unique cultures and settings where natural dynamics and human activities interact (Martin
46 et al., 2020).

47 In recent decades, globalization and new technologies have brought about rapid changes in forest
48 usage, threatening ESs in some cases (e.g. Athukorala et al., 2021). To avoid such events, it is
49 essential to understand the possible conflicts among local ES uses.

50 ESs from forests, which are major sources of ESs (Aznar-Sánchez et al., 2018), are diverse and
51 can be classified into timber and non-timber ESs. Generally, timber ESs contributes to local
52 economies and non-timber ESs secure local livelihoods (TEEB, 2010). Both types are important to a
53 local community's infrastructure and need to be continuously supplied. In addition, logging for the use
54 of timber ESs changes forest structure and may affect non-timber ES supply. Thus, timber and non-
55 timber ESs often have trade-off relationships with each other (Howe et al., 2014; Nalle et al., 2004).

56 These relationships can vary spatially across regions. For instance, among three distinct patterns
57 of timber and non-timber relationships at the forest stand scale (as depicted in Fig. 1a), foresters may
58 choose to avoid harvesting stands that have high non-timber ES values (pattern C), but relationships
59 may not always be clear (pattern B). In some cases a stand with higher non-timber ES values may also
60 supply more timber ESs (pattern A). Pattern A is likely to be the case for example for old growth
61 forests near roads which may serve as recreational areas but are also profitable for timber production.
62 These relationships give rise to regional trade-offs (as depicted in Fig. 1b) that can be viewed as
63 potential productivity frontiers and that form the basis for sustainable planning (King et al., 2015b).

64 **Fig. 1 Trade-off relationships at the forest stand scale produce different regional trade-off**
65 **relationships. Pattern A: forest stands with higher non-timber ES value are more likely to**
66 **be logged; Pattern B: there are no clear relationships between the use of timber ES and**
67 **non-timber ES values; Pattern C: foresters may avoid cutting forest stands with high non-**
68 **timber ES values.**

69 We predict that the relationship between timber and non-timber ESs can be greatly influenced by

70 environmental conditions and social context, which can differ regionally. For example, at locations
71 where rapid changes in social conditions occurred and the population is no longer aware of the
72 importance of non-timber ES, the use of the timber ESs may undermine the sustainability of non-
73 timber ESs. On the other hand, where traditional use continues or environmental awareness is high,
74 conflict among ESs may be minimized by local stakeholders. Exploring and understanding such
75 relationships in various social and environmental contexts is highly relevant for sustainable forest
76 management.

77 ESs have been modeled and mapped to identify forests that need to be conserved (Maes et al.,
78 2012; Stritih et al., 2021). Relationships such as trade-off/synergy among ES supplies have been
79 analysed by locating supply (Dai et al., 2017; Frizzle et al., 2022), demand (Adams et al., 2011; Soto
80 et al., 2018) or both (Caglayan et al., 2021; Khosravi Mashizi and Sharafatmandrad, 2021). Scenario
81 analyses have also been conducted for analysing relationships among ESs chronologically (Yamaura
82 et al., 2021). However, little research so far relates ES trade-offs to local awareness. The objectives of
83 this study were to analyse local spatial trade-off relationships between timber and non-timber ES
84 supplies in forests in Japan.

85 The use of the timber supply ES, which through logging can change forest landscapes drastically,
86 also significantly affects other ESs (Ager et al., 2017). For a sustainable supply of both timber and
87 non-timber ESs, logging locations must be selected in a manner to avoid conflicts between both types
88 of ESs. Thus, we here examine local logging tendencies in relation to the distribution of non-timber
89 ESs. We compared logging locations and ES distributions in three adjacent municipal regions in
90 Southwestern Japan with different timber harvesting practices: Saiki, Bungo-Ono, and Taketa.
91 Because distributions of non-timber ESs and logging locations differ among them, each region has
92 unique trade-off relationships. We discuss these trade-offs among ESs in relation to the environmental
93 awareness of each community as revealed in interviews with local stakeholders. This study examined
94 the hypothesis that the capability to produce timber while avoiding competition with non-timber ESs
95 is high in a region with high ES-awareness.

96 To reach our objectives, we first investigated local awareness of ESs in interviews. Then, we
97 created forest ES distribution maps from topographic and land-use maps. We evaluated three ESs

98 which are generally diminished by logging: landslide prevention, sediment retention, and forest
99 recreation. Next, we overlaid these maps with forest logging locations identified from satellite
100 imagery. Finally, we examined the observed trade-offs among ESs in relation to local attitudes to ESs.

101 **2. Materials and methods**

102 Our study area covers three adjacent municipalities, each with different geographic and social
103 conditions (see 2.1). We interviewed staff of local government and forestry enterprises in the study
104 area (see 2.2). Combining the social conditions identified in the interviews with spatial analysis of ESs
105 and logging locations, we examined the trade-offs between timber and non-timber ESs and the social
106 factors that led to them. In the spatial analysis, we divided the study area into 30-m × 30-m cells, and
107 derived non-timber ES values of each cell (see 2.3) using topography, land-use, forest type, and stand
108 age information based on the models proposed in Yamaura et al. (2021). This resulted in maps
109 indicating the value for landslide prevention, sediment retention, and forest recreation for each cell.
110 Each cell was assigned into one of four categories for each non-timber ES. These maps were overlaid
111 with logging locations identified from satellite imagery (see 2.4). In each non-timber ES class, the
112 ratios of logging area were calculated. Finally, the ratios were compared among forests with different
113 ES values to represent the local relationships among timber and non-timber ESs.

114 **2.1. Study area**

115 We studied privately owned forestry plantations in three adjacent municipal regions—Saiki, Bungo-
116 Ono, and Taketa—on Kyushu, southwestern Japan (Fig. 2), most of which are owned by individual
117 smallholders. Japanese cedar (*Cryptomeria japonica*) and hinoki cypress plantations (*Chamaecyparis*
118 *obtusa*) dominate the forested area.

119 **Fig. 2 Study area in Kyushu, southwestern Japan.**

120 Most of the plantation forests lie within 100 m of roads (Fig. 3). In Saiki, 13% of the area lies
121 more than 300 m from roads. Harvesting in such distant areas needs the construction of temporary
122 roads, which are excluded from the calculations used to create Figure 3, and are generally avoided due
123 to the high associated cost. If we consider only the slope and distance from the road, Taketa is the

124 most profitable and Saiki is the least profitable region with respect to harvesting costs.

125 Almost all forest plantations in the study area are even-aged, with unimodal age distributions
126 (Fig. 4). Clear-cutting followed by replantation is the predominant cultivation system. The relative
127 distributions of age are not significantly affected by slope or distance from roads. The median forest
128 age is >40 years, which means that there are sufficient resources to harvest forest plantations in all
129 environmental conditions in the study area. We focused on privately owned forests, where owners
130 manage their forests at their own discretion. Private forests are not subject to strong regulations on
131 logging. Local governments designate forest areas that should be preserved for disaster prevention
132 purposes, but areas can be logged if such designation is lifted. Local governments in the three regions
133 recommend clear-cutting in planted forests 40 years or older for continued timber production, and in
134 general, all timber is produced in a manner consistent with the recommendations. More than 90% of
135 the forest owners in Taketa and Bungo-Ono and about 65% in Saiki are residents within the region,
136 according to government statistics, and are therefore beneficiaries of non-timber ESs as well.

137 **Fig. 3 Distributions of forest plantations in the study area. Values are proportions of forest**
138 **plantations in the whole area of the region.**

139 **Fig. 4 Relative frequency distributions of forest plantation ages in each category of slope degree**
140 **(upper row) or distance from roads (lower row). The dashed lines are median values. Data**
141 **were drawn from the Oita Prefecture forest register.**

142 We obtained locations and age distributions of the plantation forests from the forest register
143 managed by Oita Prefecture. Terrain covariates of slope angle, aspect, and curvature were calculated
144 from a digital elevation model (DEM) provided by the GeoSpatial Information Authority of Japan
145 (GSI). The original DEM, which has a spatial resolution of 10 m, was generated through field
146 measurements and the analysis of aerial photographs. Road proximity was also calculated from the
147 DEM on a 1:25 000 digital road map provided by GSI. We obtained geology and soil type from a
148 seamless 1:200 000 digital geological map provided by the Geological Survey of Japan and a 1:50 000
149 soil map provided by GSI. All spatial data were resampled to 30-m resolution in UTM projection, and
150 slope degree and angle were calculated in SAGA v. 7.6.2 GIS software. Calculations for evaluating

151 ESs and analyzing overlays were conducted in the gdal 3.1.4, geopandas 0.9.0, pandas 1.1.5, and
152 numpy 1.16.0 packages in python v. 3.6. Model scripts used in this study are provided in the
153 supplementary material.

154 **2.2. Interviews**

155 We conducted semi-structured interviews about the current forestry situation in the study area with
156 representatives of the prefectural government, three municipal governments, and one forestry
157 enterprise in each municipality from July 2017 to December 2022. Each agency was interviewed at
158 least twice. We asked five questions: (1) who conducts the clear-cuts in the plantations in each
159 municipality, (2) attitudes towards non-timber ESs, (3) any criteria for selecting logging sites, (4) legal
160 regimes related to logging, and (5) relationships between forestry enterprises and local residents.

161 **2.3. Ecosystem service maps**

162 **2.3.1. *Landslide prevention***

163 Landslides, which occur mostly in steep areas, are restrained by tree root networks that stabilize the
164 surface soil (Stumpf and Kerle, 2011). Within and around the study area, landslides caused by heavy
165 rainfall are becoming more frequent under climate change, threatening livelihoods. The landslide
166 protection ES is therefore important in the study area, and harvesting in forests with high potential risk
167 of landslides is ideally avoided (Saito et al., 2017). Clear-cutting in susceptible forests results in
168 immature root systems and impairs the non-timber ES that prevents landslides.

169 We evaluated spatial landslide risk by using a conversion table provided by the Japan Forestry
170 Agency (JFA) to evaluate landslide hazards (MAFF, 2016). The conversion table is based on expert
171 knowledge and the Hayashi's Quantification Method-II (HQM-II), a discriminant analysis method that
172 allows qualitative variables to be used as explanatory variables (Hayashi, 1951). HQM-II categorized
173 the cells in the study area into four categories including very unstable, unstable, stable, or very stable.
174 The conversion table was created by combining expert judgement with geological factors as
175 explanatory variables and susceptibility as the objective variable. Each 30-m × 30-m forest cell was
176 scored according to geology, topography, soil depth, and forest age (Table S1) and rated as very
177 unstable, unstable, stable, or very stable on the basis of these scores (Table S2). Two kinds of geology

178 underlie most of the study area (Fig. S1): Neogene sedimentary rocks, which are generally stable, and
179 can be found in Saiki and the southern part of Bungo-Ono; and volcanic rocks, which are relatively
180 unstable, and are the most common parent material in the northern part of Bungo-Ono and Taketa.
181 Differences are also evident in the topography: the southern part of the study area has steeper and
182 more complex terrain than the northern part (Fig. S2). We estimated soil depth distributions over the
183 whole study area from soil depth and profile curvature (Table S3) according to the JFA (2011). When
184 applying the conversion table, we assumed young forests, 15 years old, to cover the entire study area.
185 These young forests are the most susceptible to landslide owing to their immature roots, In this way,
186 we excluded the influence of the current forest composition and structure and focused on the
187 theoretical landslide risk of each pixel.

188 **2.3.2. Sediment retention**

189 Soil loss not only reduces plant growth and water retention, but also causes water pollution in rivers
190 (Pimentel and Kounang, 1998). This is why sediment retention is a fundamental ES. Forest vegetation
191 moderately but surely alleviates soil loss. Tree crowns and leaf litter weaken raindrop impact, and the
192 understory vegetation holds the surface soil together (Hartanto et al., 2003). Bare soil caused by
193 logging can lead to significant soil loss. Soil can easily run off in young plantation forests where
194 understorey vegetation is sparse owing to the dense crown cover (El Kateb et al., 2013). Thus,
195 appropriate selection of logging locations is vital for conserving the sediment retention ES.

196 We estimated annual soil loss in the study area by using the Revised Universal Soil Loss
197 Equation (RUSLE: Renard et al., 1997). RUSLE estimates annual soil loss (A , $\text{t ha}^{-1} \text{ year}^{-1}$) from
198 erosion risk factors, namely rainfall-runoff erosivity (R), soil erodibility (K), slope length and
199 steepness (LS), cover and management (C), and support practice (P) (Schmidt et al., 2019), as:

$$200 \quad A = R \times K \times L \times S \times C \times P$$

201 C represents forest functions that hold the surface soil; immature forests growing in logged and
202 reforested areas are at high risk of soil erosion owing to their high C value. C can be estimated from
203 the percentage of forest floor cover (C_F , %) Miura et al. (2015) as:

$$204 \quad C = \exp(-0.051C_F).$$

205 As we did for the landslide prevention ES, we used the C_F value of a 15-year-old forest to identify

206 forests where logging should be avoided. C_F was set at 0.008 for forests with slopes less than 32
207 degrees and 0.015 for other forests based on the mean values obtained from Japanese National Forest
208 Inventory. Descriptions on the other factors are provided in the supplementary material.

209 Annual soil loss was categorized as: slight ($<1 \text{ t ha}^{-1} \text{ year}^{-1}$), moderate ($1\text{--}5 \text{ t ha}^{-1} \text{ year}^{-1}$), high
210 ($5\text{--}10 \text{ t ha}^{-1} \text{ year}^{-1}$) or significant ($>10 \text{ t ha}^{-1} \text{ year}^{-1}$). The thresholds were established with reference to
211 previous research (e.g. Fartas et al., 2022; Hagraas, 2023; Masullo, 2017).

212 2.3.3. *Forest recreation*

213 The JFA has been encouraging commercialization of the forest recreation ES for recreational uses
214 such as forest bathing (a.k.a. *shinrin-yoku*: entering forests to breathe in the clean air and to bathe in its
215 fragrance for physical and mental health) since 2019. There are high expectations of the forest
216 recreation ES as a source of tourism income to revitalize regions (Cordell et al., 2018; Starbuck et al.,
217 2006). Generally, older and larger forests are preferred for their landscape aesthetics, as an important
218 component of forest recreation (Gundersen and Frivold, 2008). This ES is higher near roads with good
219 accessibility (Abildtrup et al., 2013; Termansen et al., 2013), creating a clear trade-off with timber
220 production.

221 We evaluated the forest recreation ES on the basis of forest stand area and adjacency to roads.
222 Forest stand area is identified from the forest register, and a map of adjacency to roads is created by
223 the Accumulated Cost tool in SAGA with a slope–distance map as a cost layer created from the DEM
224 and a road distribution map provided by GSI as a source layer. Forest stand area and adjacency to
225 roads are scored according to the results of the analytical hierarchy method of Kagawa (1991, 1990)
226 which models the preferential uses of the forest for recreation. Scores were calculated as:

$$227 S_{area} = 0.0106^{exp(-3.21 \times a)}$$

$$228 S_{road} = 1 - d/100,$$

229 where S_{area} and S_{road} depend on forest stand area (a , ha) and distance from roads (d , m). We
230 multiplied these scores of each cell to calculate an index of forest recreation. Although the forest
231 recreation ES should be higher in older forests, we did not consider forest age, and instead assumed
232 that every forest in the study area is mature enough and scores depend only on forest area and road

233 proximity, because our objective was not to identify current distributions of evaluation values but to
234 estimate effects of logging on potential recreational value.

235 The values of the forest recreation ES were categorized into four classes for convenience of
236 comparison with the other two non-timber ESs: very low (< 0), low (0–0.2), high (0.2–0.35) and very
237 high (> 0.35). These thresholds were determined so that the evaluated cell areas categorized as low,
238 high and very high were approximately the same.

239 **2.4. Logging locations**

240 Logging locations were identified from satellite imagery. We used a forest disturbance dataset for all
241 of Japan (Shimizu and Saito, 2021). This dataset was generated by using a pixel-level Landsat time-
242 series analysis and a random forest classification approach to map annual forest disturbance types (i.e.
243 logging, conversion, thinning and natural disturbances) at a 30-m spatial resolution from 1985 to 2019.
244 For the logging class, the dataset has a producer accuracy of 80.1% and a user accuracy of 93.8%. We
245 clipped the dataset to the study area and used logging locations detected from 2015 to 2019 for this
246 study.

247 To calculate the logged area ratio in each ES category, numbers of logged/not logged raster cells
248 were counted and summarized for each ES category. The logged/not logged area (ha) was calculated
249 by multiplying the number of raster cells by 0.09, the area of each raster cell, and the logged area ratio
250 was calculated by dividing the number of logged records by that of all raster cells in each category.
251 These calculated logged area ratios were then subjected to comparative analysis, and the statistical
252 significance was elucidated through a two-tailed test ($P < 0.1$) for each municipality.

253 **3. Results**

254 **3.1. Interviews**

255 Across all regions, logging locations were predominantly selected by forest owners' associations or
256 forestry enterprises, who then approached forest owners about the potential for logging. Typically,
257 forest owners have limited interest in the management of their forests and rely on the suggestions of
258 the forestry enterprises. Although forestry enterprises are obligated to inform local government where

259 they are logging, local government lacks the authority to control logging.

260 Saiki produces more timber than the other two regions. The forest owners' association plays a
261 central role in timber supply there (Table 1). Most forests in Saiki cover steep slopes, and local people
262 recognize the need for protecting mountainous lands. In addition, since Saiki faces the sea and has a
263 thriving fishing industry, they want to avoid water pollution caused by soil loss. Thus, awareness of
264 non-timber ESs is high in this region. The forest owners' association in Saiki sets criteria for selecting
265 logging sites based on their experience to avoid harvestings in forests where landslides and soil loss
266 occur easily.

267 A biomass power plant in Bungo-Ono commenced operations in 2016, but local forest owners
268 had been amassing timber for energetic use already several years before. Although statistical data are
269 lacking, local governments have observed a significant surge in logging activities by non-regional
270 forestry companies, corresponding with rapidly increasing demand for timber. The municipal
271 government stated that these enterprises have limited engagement with local residents, leading to
272 possible concerns of limited corporate responsibility towards the local ESs.

273 In Taketa, several local companies conduct small-scale logging operations and are responsible for
274 most of the logging in the region. The companies are fond of maintaining a good reputation in the
275 local community and hence try to avoid ES degradation caused by their logging. However, they have
276 not established any specific criteria to conserve ESs in the selection of logging sites.

277 **Table. 1 Comparison of logging practices and environmental awareness in the three regions.**

278 Thus, landslide prevention and soil retention are critical concerns in Saiki and Taketa. Contrarily,
279 the forest recreation ES is of no interest in any region. There was only provisional awareness of this
280 ES, which is a new concept among residents, and little attention was paid to it.

281 **3.2. Ecosystem service maps**

282 The distribution of forests with a high risk for landslides is similar to that of steep slopes (Figs. 5a,
283 S2), with lower risks in forested areas underlain by Neogene sedimentary rocks (Fig. S1). Taketa has
284 the largest area of high-risk forests, and Saiki the lowest. Although Saiki has a wide distribution of
285 steep terrain, the risk of landslide is lower because of the sedimentary geology (Fig. S1). Soil loss

286 shows the opposite results to landslide, with Saiki having the highest risk and Taketa the lowest (Fig.
287 5b). In Saiki, Brown Forest soils, which are easily eroded, cover steep slopes, posing high risk and the
288 need for careful consideration in the selection of logging sites (Fig. S3). On the other hand, Taketa
289 was evaluated as having less risk because it is covered with Black soil, which is relatively stable on
290 moderate slopes. The forest recreation ES map shows that Saiki has the largest low-value area (Fig.
291 5c), most likely caused by the low road network density due to the steepness of the terrain. Bungo-
292 Ono and Taketa have similar distributions of this ES.

293 **Fig. 5 Distributions of ecosystem service values of (a) landslide susceptibility, (b) sediment**
294 **retention and (c) forest recreation.**

295 **3.3. Logging locations**

296 The logged area ratio in Saiki is about twice of that observed in the other two regions (Table 2).
297 Logging is evenly distributed throughout most of Saiki, although it is sparse along the eastern coast
298 (Fig. 6). The area ratios of logged forests were higher on gentle slopes, where both logging cost and
299 hazard risk are low (Fig. 7). On the other hand, forests near roads, where the forest recreation ES is
300 high, are logged more frequently. In Bungo-Ono, logging is relatively common in the southern part.
301 Forests on steep slopes were also logged frequently there (Fig. 7). In Taketa, logged forests were
302 distributed mainly in the northern and southern parts, and to a minor extent in the central part.

303 **Fig. 6 Locations of forests logged in 2015–2019.**

304 **Fig. 7 Logged area ratio in each category. Each value shows the ratio of logged area to forest**
305 **plantations in each category (sum = 1.0).**

306 **Table. 2 Logged area ratio in each region in 2015–2019.**

307 **3.4. Overlaying maps**

308 Figure 8 shows the relationships between the ESs and logging locations in each municipality. The
309 ratios of logged area show that the spatial relationships with ESs differ substantially. In Saiki, the ratio
310 of logged area is significantly smaller in landslide susceptibility classes of very unstable and unstable
311 forests than in stable and very stable. However, the overall logged area is large, resulting in a large

312 logged area in the high-risk category. In Taketa, which has the greatest area susceptible to landslide
313 (Fig. 5a), the ratios of logging area were lower in the two least stable classes than in the two most
314 stable classes. In Bungo-Ono, the ratio was slightly but significantly lower in stable forests than in
315 very stable, unstable, and very unstable forests, which did not differ. Foresters from outside of the
316 municipality might have harvested timber even in forests at high risk owing to lack of knowledge or
317 concern with regard to landslide risk.

318 **Fig. 8 Logged area ratio of each category with (I) landslide susceptibility class (x-axis) a, very**
319 **stable; b, stable; c, unstable; d, very unstable; (II) annual soil loss class (x-axis) a, slight (<1**
320 **t ha⁻¹ year⁻¹); b, moderate (1–5); c, high (5–10); d, significant (>10); (III) forest recreation**
321 **index class (x-axis): a, low (0); b, moderate (0–0.2); c, high (0.2–0.35); d, very high (>0.35).**
322 **Letters on the red curves indicate other classes from which each class differed significantly**
323 **($P < 0.1$).**

324 Saiki had the largest area in the category at the highest risk of soil loss, and Taketa had the
325 smallest (Fig. 8II). In Saiki, the logged area ratio was significantly lower in the category at highest risk
326 of soil loss than in the categories at lower risk. The category at lowest risk might have insufficient
327 forested area to detect significance. In Taketa, the logged area ratio was significantly higher in the two
328 lower risk categories than in the two higher risk categories. These spatial logging tendencies
329 contribute to the maintenance of the sediment retention ES, especially in Saiki, where there is a
330 generally high risk of soil loss. In Bungo-Ono, on the other hand, the logged area ratio was higher in
331 unstable forests. As regards the landslide prevention ES, this spatial relationship could lead to ES
332 degradation.

333 The relationships between logging locations and the forest recreation ES were obscure (Fig. 8III).
334 In each municipality, the forested area in the lowest rated category was the largest. Distributions were
335 similar among them, but the forest area ratio with the lowest values was largest in Saiki. There was no
336 overall consistent tendency in logging ratios, although there were significant differences. In Taketa
337 only, the logged ratio was highest in the category with the lowest forest recreation ES, even though
338 these forests are disadvantageous for logging operations owing to long distances from roads.

339 According to the relative frequency distribution of forest ages in each ES category (Fig. 9),
340 although there are slight differences in age distributions among categories, the median values in all
341 categories exceed the recommended harvest age of 40 years. Figure 9 suggests that resources are
342 sufficient and do not constrain wood production in any category.

343 **Fig. 9 Relative frequency distribution of forest plantation ages in each category of landslide**
344 **susceptibility class (upper row): a, very stable; b, stable; c, unstable; d, very unstable;**
345 **annual soil loss class (middle row): a, slight ($<0.1 \text{ t ha}^{-1} \text{ year}^{-1}$); b, moderate (0.1–1); c,**
346 **high (1–10); d, significant (>10); or forest recreation class (lower row) : a, <0.1 ; b, 0.1–0.25;**
347 **c, 0.25–0.5; d, ≥ 0.5 . The dashed lines are median values.**

348 4. Discussion

349 We explored trade-off relationships between timber and non-timber ESs in three adjacent
350 municipalities and found great variations among them. In regions where local timber production is
351 closely tied to the local community, as in Saiki and Taketa, spatial conflicts between timber and non-
352 timber ESs tended to be lower. According to the foresters in Saiki and Taketa, logging is usually
353 avoided in areas with important non-timber ESs, as these benefit local residents and should be
354 conserved. The foresters also expressed fear of damaging their reputation in the community. Thus,
355 foresters seem to avoid logging areas with high local ES values (pattern C in Fig. 1a). In contrast,
356 logging locations in Bungo-Ono, which were selected by external forestry enterprises, were not
357 systematically related with non-timber ES distributions. These forest enterprises have a tenuous
358 relationship with local residents. According to the municipal officer, foresters from outside the
359 municipality have been crossing and logging along the southern boundary, and the fact that biomass
360 power plants have started operating as a destination for timber consumption is likely to have fueled
361 this trend. At the same time, a greater area of forests in the southern part is at high risk of landslide
362 and soil loss (Fig. 5a). As a result, the logging ratio is not low even in forests at high risk (Fig. 8). This
363 may lead to regional conflicts between timber and non-timber ESs, negatively affecting local
364 sustainability (Fig. 1b).

365 **4.1. Relationships between non-timber ecosystem services and logging locations**

366 As the results indicate, local environmental awareness positively affects ES trade-offs. Sometimes,
367 lack of awareness may reduce ESs, as in Bungo-Ono, where the social context has rapidly changed. In
368 Saiki and Taketa, in contrast, the traditional use of ESs has had the opposite effect. Thus, our
369 hypotheses that conflicts among ESs are moderate in regions with high ES-awareness are generally
370 confirmed. Local awareness is shaped by local history and culture (Chen et al., 2019). However, in
371 today's global society, numerous regions are experiencing major changes in the use of ESs
372 (Brown, 2013; Jansson et al., 2015). In such regions, a different ES trade-off from before might
373 emerge and cause conflict (King et al., 2015a). Research examining how changes in social conditions
374 affect ESs and the consequences for our lives will become even more important (Carpenter et al.,
375 2009; Chapin et al., 2010). An important research gap to be addressed is the relationship between
376 environmental awareness and ES trade-offs. Quantifying environmental awareness by inquiry
377 (Schuman and Presser, 1996) or by surveys on willingness to pay (Nelson et al., 2008) may enable
378 such relationships to be statistically analysed.

379 We evaluated and mapped ESs in forest plantations that supply various ESs in addition to timber.
380 Attention to such ESs has been rising, and earlier studies identified a great influence of terrain factors
381 (e.g. Jackson et al., 2013; Yamaura et al., 2021). Our evaluation maps depended significantly on
382 topography but also on geology, as shown clearly by the comparison among regions. For example,
383 landslide susceptibility is lower in Saiki, despite its steep slopes, on account of the underlying
384 geology. Thus, a comprehensive evaluation of geography forms an important basis in mapping the
385 distribution of ESs. Although the topography in Saiki seems to be economically disadvantageous for
386 forestry (Fig. 3), our results indicate that the municipality is well suited to logging owing to its high
387 soil stability. In fact, timber production is more prevalent in this region than in other areas. While
388 previous studies have reported that variables related to the profitability of a forestry operation, such as
389 slope angle and distance from the road, affect the probability of harvesting (Beach et al., 2005;
390 Polyakov et al., 2010; Prestemon and Wear, 2000), our study suggests that there may also be a
391 relationship between ESs and harvesting. Identifying the factors involved in the selection of logging

392 sites will contribute to the development of efficient policies for protecting ecosystem services from
393 logging (Yamada, 2020).

394 In all municipalities, relationships between the forest recreation ES and logging locations were
395 unclear (Fig. 8). Road proximity, a key factor in the forest recreation value, did not affect logged ratios
396 in any region (Fig. 7). Even though forests with high forest recreation value are profitable for logging,
397 there seem to be no critical conflicts. The forest recreation ES has not been realized by local residents
398 as a benefit of forest ecosystems. The interviews revealed that the residents consider the forests to be
399 the backdrop to their daily lives, and forest recreation to be irrelevant. Ninan and Inoue (2013)
400 estimated forest recreation benefits at about USD 140 to 145 per hectare per year in other regions of
401 Japan. Forest tourism also has a positive effect on local economic development in other countries
402 (Archabald and Naughton-Treves, 2001; Mayer, 2014). However, spending related to forest
403 recreation, including camping and guide fees, accounts for less than 1% of the tourism industry in the
404 region (Oita Prefectural Government, 2013). Stimulating economic activities using the forest
405 recreation ES can be expected to increase local awareness (Brandt and Buckley, 2018), and would
406 therefore help to ease the competitive relationship with timber production as in other non-timber ESs
407 observed in this research.

408 **4.2. Efficient forest management methods for the study area**

409 On the basis of our results, we suggest guidelines for avoiding conflicts among ESs that are
410 appropriate for local logging practices. Public awareness of ESs is high in Saiki and Taketa. It would
411 hence be beneficial to assist, but not regulate, foresters in selecting appropriate logging sites by
412 offering ES information based on scientific analysis, e.g., ES evaluation maps (Maes et al., 2012;
413 Schägner et al., 2013). In Bungo-Ono, where ESs seem to be of less concern, it would be beneficial to
414 identify foresters responsible for logging operations in the region and focus on raising their awareness
415 of ESs. Regulation of logging sites (Yamada and Yamaura, 2017) and direct payment schemes
416 (Kemkes et al., 2010; Polasky et al., 2014) may also be effective tools.

417 Although the three regions in our study are adjacent, ES relationships differ among them,
418 highlighting the importance of considering local policy-making. A top-down approach, by which

419 policies are formulated in a limited number of people, sometimes ignoring the local population, may
420 not only provoke backlash from residents but also upset the existing equilibrium and impair ESs
421 (Fraser et al., 2006). A bottom-up approach, in which local stakeholders participate in establishing
422 policy, is required instead, since the policy will reflect each side's requirements for how the local ESs
423 will be used (Ananda, 2007; Kangas et al., 1996). Moreover, this approach should raise participants'
424 environmental awareness (Fraser et al., 2006). However, not all stakeholders are familiar enough with
425 the current status and trade-offs of ESs in their local forests. Providing information based on analyses
426 such as the distribution and usage tendencies of ESs in a region is essential in a bottom-up approach
427 (Evans et al., 2018). As this study shows, we must first understand local harvesting practices, how ESs
428 are used (and the effect of social context) and how their interactions have been shaped in a region.
429 Presented with this information, local committees can establish local forest management policies to
430 meet local demands for ESs (Maes et al., 2012).

431 **5. Conclusion**

432 To manage local forests sustainably, clear-cutting should be avoided in forests where a supply of
433 high-value non-timber ESs is expected. Overlaying ES and logging location maps revealed the
434 relationships between timber and non-timber ESs. Saiki where a forest owners' association manages
435 the logging operations of the whole region, showed an ES-friendly spatial logging tendency. On the
436 other hand, as shown in Bungo-Ono, non-timber ESs could be damaged if non-local foresters who are
437 unaware of local conditions are responsible for logging. Thus, local awareness and the relationships
438 between timber and non-timber ESs differ among regions. Sustainable provision of ESs, therefore,
439 requires localized rather than uniform policymaking, and we conclude that exploring local awareness
440 and relationships among ESs is the first step towards developing sustainable forest management
441 strategies. For example, information based on scientific analysis would be helpful where local
442 residents have a high awareness of ESs. On the other hand, education is required to make uninformed
443 foresters aware of ESs in regions where forests with high non-timber ES values are being logged.

444 **Data availability statement**

445 The geospatial data underlying this article are available online at <https://www.gsi.go.jp>. Logging
446 location data are published in Zenodo at <https://doi.org/10.5281/zenodo.4654619>.

447 **Supplementary material**

448 Supplementary data are available at *Forestry* online.

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453 **Conflict of interest statement**

454 The authors declare that they have no known competing financial interests or personal relationships
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459 **References**

- 460 Abildtrup, J., Garcia, S., Olsen, S. B., and Stenger, A. 2013 Spatial preference heterogeneity in forest
461 recreation, *Ecol. Econ.*, **92**, 67–77. <https://doi.org/10.1016/J.ECOLECON.2013.01.001>
- 462 Adams, D. C., Bwenge, A. N., Lee, D. J., Larkin, S. L., and Alavalapati, J. R. R. 2011 Public
463 preferences for controlling upland invasive plants in state parks: Application of a choice model,
464 *For. Policy Econ.*, **13**(6), 465–472. <https://doi.org/10.1016/J.FORPOL.2011.04.003>
- 465 Ager, A. A., Vogler, K. C., Day, M. A., and Bailey, J. D. 2017 Economic Opportunities and Trade-
466 Offs in Collaborative Forest Landscape Restoration, *Ecol. Econ.*, **136**, 226–239.
467 <https://doi.org/10.1016/J.ECOLECON.2017.01.001>
- 468 Ananda, J. 2007 Implementing Participatory Decision Making in Forest Planning, *Environ. Manage.*,

469 39(4), 534–544. <https://doi.org/10.1007/s00267-006-0031-2>

470 Archabald, K., and Naughton-Treves, L. 2001 Tourism revenue-sharing around national parks in
471 Western Uganda: early efforts to identify and reward local communities, *Environ. Conserv.*,
472 28(2), 135–149. <https://doi.org/10.1017/S0376892901000145>

473 Asah, S. T., Guerry, A. D., Blahna, D. J., and Lawler, J. J. 2014 Perception, acquisition and use of
474 ecosystem services: Human behavior, and ecosystem management and policy implications,
475 *Ecosyst. Serv.*, **10**, 180–186. <https://doi.org/10.1016/J.ECOSER.2014.08.003>

476 Athukorala, D., Estoque, R. C., Murayama, Y., and Matsushita, B. 2021 Ecosystem Services
477 Monitoring in the Muthurajawela Marsh and Negombo Lagoon, Sri Lanka, for Sustainable
478 Landscape Planning, *Sustainability*, **13**(20), 11463. <https://doi.org/10.3390/su132011463>

479 Aznar-Sánchez, J., Belmonte-Ureña, L., López-Serrano, M., Velasco-Muñoz, J., Aznar-Sánchez, J. A.,
480 Belmonte-Ureña, L. J., López-Serrano, M. J., and Velasco-Muñoz, J. F. 2018 Forest Ecosystem
481 Services: An Analysis of Worldwide Research, *Forests*, **9**(8), 453.
482 <https://doi.org/10.3390/f9080453>

483 Barlow, J., Lennox, G. D., Ferreira, J., Berenguer, E., Lees, A. C., MacNally, R., Thomson, J. R.,
484 Ferraz, S. F. de B., Louzada, J., Oliveira, V. H. F., Parry, L., Ribeiro de Castro Solar, R., Vieira,
485 I. C. G., Aragão, L. E. O. C., Begotti, R. A., Braga, R. F., Cardoso, T. M., Jr, R. C. de O., Souza
486 Jr, C. M., Moura, N. G., Nunes, S. S., Siqueira, J. V., Pardini, R., Silveira, J. M., Vaz-de-Mello,
487 F. Z., Veiga, R. C. S., Venturieri, A., and Gardner, T. A. 2016 Anthropogenic disturbance in
488 tropical forests can double biodiversity loss from deforestation, *Nature*, **535**(7610), 144–147.
489 <https://doi.org/10.1038/nature18326>

490 Beach, R. H., Pattanayak, S. K., Yang, J.-C., Murray, B. C., and Abt, R. C. 2005 Econometric studies
491 of non-industrial private forest management: a review and synthesis, *For. Policy Econ.*, **7**(3),
492 261–281. [https://doi.org/10.1016/S1389-9341\(03\)00065-0](https://doi.org/10.1016/S1389-9341(03)00065-0)

493 Brandt, J. S., and Buckley, R. C. 2018 A global systematic review of empirical evidence of ecotourism
494 impacts on forests in biodiversity hotspots, *Curr. Opin. Environ. Sustain.*, **32**, 112–118.
495 <https://doi.org/10.1016/J.COSUST.2018.04.004>

496 Brown, G. 2013 The relationship between social values for ecosystem services and global land cover:

497 An empirical analysis, *Ecosyst. Serv.*, **5**, 58–68. <https://doi.org/10.1016/j.ecoser.2013.06.004>

498 Caglayan, İ., Yeşil, A., Kabak, Ö., and Bettinger, P. 2021 A decision making approach for assignment
499 of ecosystem services to forest management units: A case study in northwest Turkey, *Ecol.*
500 *Indic.*, **121**, 107056. <https://doi.org/10.1016/j.ecolind.2020.107056>

501 Carpenter, S. R., Mooney, H. A., Agard, J., Capistrano, D., Defries, R. S., Diaz, S., Dietz, T.,
502 Duraiappah, A. K., Oteng-Yeboah, A., Pereira, H. M., Perrings, C., Reid, W. V., Sarukhan, J.,
503 Scholes, R. J., and Whyte, A. 2009 Science for managing ecosystem services: Beyond the
504 Millennium Ecosystem Assessment, *Proc. Natl. Acad. Sci.*, **106**(5), 1305–1312.
505 <https://doi.org/10.1073/PNAS.0808772106>

506 Chapin, F. S., Carpenter, S. R., Kofinas, G. P., Folke, C., Abel, N., Clark, W. C., Olsson, P., Smith, D.
507 M. S., Walker, B., Young, O. R., Berkes, F., Biggs, R., Grove, J. M., Naylor, R. L., Pinkerton,
508 E., Steffen, W., and Swanson, F. J. 2010 Ecosystem stewardship: sustainability strategies for a
509 rapidly changing planet, *Trends Ecol. Evol.*, **25**(4), 241–249.
510 <https://doi.org/10.1016/j.tree.2009.10.008>

511 Chen, J., Jiang, B., Bai, Y., Xu, X., and Alatalo, J. M. 2019 Quantifying ecosystem services supply
512 and demand shortfalls and mismatches for management optimisation, *Sci. Total Environ.*, **650**,
513 1426–1439. <https://doi.org/10.1016/J.SCITOTENV.2018.09.126>

514 Cordell, H. K., Bergstrom, J. C., and Watson, A. E. 2018 Economic Growth and Interdependence
515 Effects of State Park Visitation in Local and State Economies,
516 <https://doi.org/10.1080/00222216.1992.11969892>, **24**(3), 253–268.
517 <https://doi.org/10.1080/00222216.1992.11969892>

518 Dai, E. F., Wang, X. L., Zhu, J. J., and Xi, W. M. 2017 Quantifying ecosystem service trade-offs for
519 plantation forest management to benefit provisioning and regulating services, *Ecol. Evol.*, **7**(19),
520 7807–7821. <https://doi.org/10.1002/ECE3.3286>

521 Dalimunthe, S. 2018 Who Manages Space? Eco-DRR and the Local Community, *Sustainability*,
522 **10**(6), 1705. <https://doi.org/10.3390/su10061705>

523 El Kateb, H., Zhang, H., Zhang, P., and Mosandl, R. 2013 Soil erosion and surface runoff on different
524 vegetation covers and slope gradients: A field experiment in Southern Shaanxi Province, China,

525 *CATENA*, **105**, 1–10. <https://doi.org/10.1016/j.catena.2012.12.012>

526 Evans, K., Guariguata, M. R., and Brancalion, P. H. S. 2018 Participatory monitoring to connect local
527 and global priorities for forest restoration, *Conserv. Biol.*, **32**(3), 525–534.
528 <https://doi.org/10.1111/cobi.13110>

529 Fartas, N., Fellah, B. El, Mastere, M., Benzougagh, B., and Brahim, M. El 2022 Potential Soil
530 Erosion Modeled with RUSLE Approach and Geospatial Techniques (GIS Tools and Remote
531 Sensing) in Oued Joumouaa Watershed (Western Prerif-Morocco), *Iraqi Geol. J.*, **55**(2), 47–61.
532 <https://doi.org/10.46717/IGJ.55.2B.5MS-2022-08-21>

533 Fraser, E. D. G., Dougill, A. J., Mabee, W. E., Reed, M., and McAlpine, P. 2006 Bottom up and top
534 down: Analysis of participatory processes for sustainability indicator identification as a pathway
535 to community empowerment and sustainable environmental management, *J. Environ. Manage.*,
536 **78**(2), 114–127. <https://doi.org/10.1016/j.jenvman.2005.04.009>

537 Frizzle, C., Fournier, R. A., Trudel, M., and Luther, J. E. 2022 Towards sustainable forestry: Using a
538 spatial Bayesian belief network to quantify trade-offs among forest-related ecosystem services, *J.*
539 *Environ. Manage.*, **301**, 113817. <https://doi.org/10.1016/J.JENVMAN.2021.113817>

540 Gardner, T. A., Barlow, J., Sodhi, N. S., and Peres, C. A. 2010 A multi-region assessment of tropical
541 forest biodiversity in a human-modified world, *Biol. Conserv.*, **143**(10), 2293–2300.
542 <https://doi.org/10.1016/j.biocon.2010.05.017>

543 Gundersen, V. S., and Frivold, L. H. 2008 Public preferences for forest structures: A review of
544 quantitative surveys from Finland, Norway and Sweden, *Urban For. Urban Green.*, **7**(4), 241–
545 258. <https://doi.org/10.1016/J.UFUG.2008.05.001>

546 Hagra, A. 2023 Estimating water erosion in the EL-Mador Valley Basin, South-West Matrouh City,
547 Egypt, using revised universal soil loss equation (RUSLE) model through GIS, *Environ. Earth*
548 *Sci.*, **82**(1), 1–17. <https://doi.org/10.1007/S12665-022-10722-0/TABLES/4>

549 Hartanto, H., Prabhu, R., Widayat, A. S. ., and Asdak, C. 2003 Factors affecting runoff and soil
550 erosion: plot-level soil loss monitoring for assessing sustainability of forest management, *For.*
551 *Ecol. Manage.*, **180**(1), 361–374. [https://doi.org/10.1016/S0378-1127\(02\)00656-4](https://doi.org/10.1016/S0378-1127(02)00656-4)

552 Hayashi, C. 1951 On the prediction of phenomena from qualitative data and the quantification of

553 qualitative data from the mathematico-statistical point of view, *Ann. Inst. Stat. Math.*, **3**(1), 69–
554 98. <https://doi.org/10.1007/BF02949778/METRICS>

555 Howe, C., Suich, H., Vira, B., and Mace, G. M. 2014 Creating win-wins from trade-offs? Ecosystem
556 services for human well-being: A meta-analysis of ecosystem service trade-offs and synergies in
557 the real world, *Glob. Environ. Chang.*, **28**(1), 263–275.
558 <https://doi.org/10.1016/J.GLOENVCHA.2014.07.005>

559 Jackson, B., Pagella, T., Sinclair, F., Orellana, B., Henshaw, A., Reynolds, B., McIntyre, N., Wheeler,
560 H., and Eycott, A. 2013 Polyscape: A GIS mapping framework providing efficient and spatially
561 explicit landscape-scale valuation of multiple ecosystem services, *Landsc. Urban Plan.*, **112**(1),
562 74–88. <https://doi.org/10.1016/J.LANDURBPLAN.2012.12.014>

563 Jansson, R., Nilsson, C., Keskitalo, E. C. H., Vlasova, T., Sutinen, M.-L., Moen, J., Chapin, F.,
564 Bråthen, K., Cabeza, M., Callaghan, T. V, Oort, B. Van, Dannevig, H., Bay-Larsen, I. A., Ims, R.
565 A., Jansson, R., Nilsson, C., Carina, E., Keskitalo, H., Vlasova, T., Sutinen, M.-L., Moen, J.,
566 Stuart, F., Iii, C., Bråthen, K. A., Callaghan, T. V, Van Oort, B., Dannevig, H., Bay-Larsen 13, I.
567 A., Ims, R. A., and Aspholm, P. E. 2015 Future changes in the supply of goods and services from
568 natural ecosystems: prospects for the European north, *Ecol. Soc. Publ. Online Sep 14, 2015* |
569 *Doi10.5751/ES-07607-200332*, **20**(3). <https://doi.org/10.5751/ES-07607-200332>

570 JFA (Japan Forestry Agency) 2011 *Report on Survey on Countermeasures to Control Radioactive*
571 *Substances Discharge in Forested Areas, Tokyo. in Japanese.*

572 Kagawa, T. 1990 A Study on the Amenity of Kitayama Artificial Forest, *J. Japanese Inst. Landsc.*
573 *Archit.*, **54**(5), 185–190. https://doi.org/10.5632/jila1934.54.5_185

574 Kagawa, T. 1991 Studies on the Amenity of Coppice and Natural Forest, *J. Japanese Inst. Landsc.*
575 *Archit.*, **55**(5), 217–222. https://doi.org/10.5632/jila1934.55.5_217

576 Kangas, J., Loikkanen, T., Pukkala, T., and Pykäläinen, J. 1996 A participatory approach to tactical
577 forest planning., *Acta For. Fenn.*, retrieved December 6, 2016 from internet:
578 <https://helda.helsinki.fi/handle/10138/27291>, **251**, 1–24.

579 Karjala, M. K., Karjala, M. K., Sherry, E. E., Sherry, E. E., Dewhurst, S. M., and Dewhurst, S. M.
580 2004 Criteria and indicators for sustainable forest planning: a framework for recording

581 Aboriginal resource and social values, *For. Policy Econ.*, **6**, 95–110.
582 <https://doi.org/10.1016/S1389-9341>

583 Katila, M., Rajala, T., and Kangas, A. 2020 Assessing local trends in indicators of ecosystem services
584 with a time series of forest resource maps, *Silva Fenn.*, **54**(4). <https://doi.org/10.14214/sf.10347>

585 Kemkes, R. J., Farley, J., and Koliba, C. J. 2010 Determining when payments are an effective policy
586 approach to ecosystem service provision, *Ecol. Econ.*, **69**(11), 2069–2074.
587 <https://doi.org/10.1016/J.ECOLECON.2009.11.032>

588 Khosravi Mashizi, A., and Sharafatmandrad, M. 2021 Investigating tradeoffs between supply, use and
589 demand of ecosystem services and their effective drivers for sustainable environmental
590 management, *J. Environ. Manage.*, **289**, 112534.
591 <https://doi.org/10.1016/J.JENVMAN.2021.112534>

592 King, E., Cavender-Bares, J., Balvanera, P., Mwampamba, T. H., and Polasky, S. 2015a Trade-offs in
593 ecosystem services and varying stakeholder preferences: evaluating conflicts, obstacles, and
594 opportunities, *Ecol. Soc.*, **20**(3). <https://doi.org/10.5751/ES-07822-200325>

595 King, S. L., Schick, R. S., Donovan, C., Booth, C. G., Burgman, M., Thomas, L., and Harwood, J.
596 2015b An interim framework for assessing the population consequences of disturbance, *Methods*
597 *Ecol. Evol.*, **6**(10), 1150–1158. <https://doi.org/10.1111/2041-210X.12411>

598 Maes, J., Egoh, B., Willemsen, L., Liqueste, C., Vihervaara, P., Schägner, J. P., Grizzetti, B., Drakou, E.
599 G., Notte, A. La, Zulian, G., Bouraoui, F., Luisa Paracchini, M., Braat, L., and Bidoglio, G. 2012
600 Mapping ecosystem services for policy support and decision making in the European Union,
601 *Ecosyst. Serv.*, **1**(1), 31–39. <https://doi.org/10.1016/J.ECOSER.2012.06.004>

602 MAFF 2016 *Investigation method of the disaster risk district in the mountain*, Tokyo.

603 Martin, D. A., Osen, K., Grass, I., Hölscher, D., Tschardtke, T., Wurz, A., and Kreft, H. 2020 Land-
604 use history determines ecosystem services and conservation value in tropical agroforestry,
605 *Conserv. Lett.*, **13**(5), e12740. <https://doi.org/10.1111/CONL.12740>

606 Masullo, A. 2017 Organic wastes management in a circular economy approach: Rebuilding the link
607 between urban and rural areas, *Ecol. Eng.*, **101**, 84–90.
608 <https://doi.org/10.1016/J.ECOLENG.2017.01.005>

609 Mayer, M. 2014 Can nature-based tourism benefits compensate for the costs of national parks? A
610 study of the Bavarian Forest National Park, Germany, *J. Sustain. Tour.*, **22**(4), 561–583.
611 <https://doi.org/10.1080/09669582.2013.871020>

612 Millenium Ecosystem Assessment 2005 *Ecosystems and Human Well-Being: Synthesis.*, Island Press,
613 Washinton, DC.

614 Miura, S., Ugawa, S., Yoshinaga, S., Yamada, T., and Hirai, K. 2015 Floor cover percentage
615 determines splash erosion in chamaecyparis obtusa forests, *Soil Sci. Soc. Am. J.*, **79**(6), 1782–
616 1791. <https://doi.org/10.2136/sssaj2015.05.0171>

617 Nalle, D. J., Montgomery, C. A., Arthur, J. L., Polasky, S., and Schumaker, N. H. 2004 Modeling joint
618 production of wildlife and timber, *J. Environ. Econ. Manage.*, **48**(3), 997–1017.
619 <https://doi.org/10.1016/J.JEEM.2004.01.001>

620 Nelson, E., Polasky, S., Lewis, D. J., Plantinga, A. J., Lonsdorf, E., White, D., Bael, D., and Lawler, J.
621 J. 2008 Efficiency of incentives to jointly increase carbon sequestration and species conservation
622 on a landscape, *Proc. Natl. Acad. Sci.*, **105**(28), 9471–9476.
623 <https://doi.org/10.1073/PNAS.0706178105>

624 Ninan, K. N., and Inoue, M. 2013 Valuing forest ecosystem services: Case study of a forest reserve in
625 Japan, *Ecosyst. Serv.*, **5**, 78–87. <https://doi.org/10.1016/j.ecoser.2013.02.006>

626 Oita Prefectural Government 2013 *The economic ripple effect of tourists' and visitors' consumption*
627 *on the economic ripple effects on the prefecture's industries*, Oita, pp. 30. in *Japanese*

628 Orenstein, D. E., and Groner, E. 2014 In the eye of the stakeholder: Changes in perceptions of
629 ecosystem services across an international border, *Ecosyst. Serv.*, **8**, 185–196.
630 <https://doi.org/10.1016/J.ECOSER.2014.04.004>

631 Pimentel, D., and Kounang, N. 1998 Ecology of Soil Erosion in Ecosystems, *Ecosystems*, **1**(5), 416–
632 426. <https://doi.org/10.1007/s100219900035>

633 Polasky, S., Lewis, D. J., Plantinga, A. J., and Nelson, E. 2014 Implementing the optimal provision of
634 ecosystem services, *Proc. Natl. Acad. Sci.*, **111**(17). <https://doi.org/10.1073/pnas.1404484111>

635 Polyakov, M., Wear, D. N., and Huggett, R. N. 2010 Harvest Choice and Timber Supply Models for
636 Forest Forecasting, *For. Sci.*, **56**(4), 344–355.

637 <https://doi.org/10.1093/FORRESTSCIENCE/56.4.344>
638
639 Prestemon, J. P., and Wear, D. N. 2000 Linking Harvest Choices to Timber Supply, *For. Sci.*, **46**(3),
640 377–389. <https://doi.org/10.1093/FORRESTSCIENCE/46.3.377>
641 Renard, K. G., Foster, G. R., Weesies, G. A., McCool, D. K., and Yoder, D. C. 1997 *Predicting Soil*
642 *Erosion by Water: A Guide to Conservation Planning with the Revised Universal Soil Loss*
643 *Equation (RUSLE). Agriculture Handbook No. 703.*, U.S. Department of Agriculture,
644 Agricultural Research Service, Washington, District of Columbia, USA, pp. 404.
645 Saito, H., Murakami, W., Daimaru, H., and Oguchi, T. 2017 Effect of forest clear-cutting on landslide
646 occurrences: Analysis of rainfall thresholds at Mt. Ichifusa, Japan, *Geomorphology*, **276**, 1–7.
647 <https://doi.org/10.1016/J.GEOMORPH.2016.09.024>
648 Santika, T., Wilson, K. A., Budiharta, S., Kusworo, A., Meijaard, E., Law, E. A., Friedman, R.,
649 Hutabarat, J. A., Indrawan, T. P., St. John, F. A. V., and Struebig, M. J. 2019 Heterogeneous
650 impacts of community forestry on forest conservation and poverty alleviation: Evidence from
651 Indonesia, *People Nat.*, **1**(2), pan3.25. <https://doi.org/10.1002/pan3.25>
652 Schägner, J. P., Brander, L., Maes, J., and Hartje, V. (June 1, 2013): Mapping ecosystem services'
653 values: Current practice and future prospects, *Ecosyst. Serv.*, Elsevier.
654 <https://doi.org/10.1016/j.ecoser.2013.02.003>
655 Schmidt, S., Tresch, S., and Meusburger, K. 2019 Modification of the RUSLE slope length and
656 steepness factor (LS-factor) based on rainfall experiments at steep alpine grasslands, *MethodsX*,
657 **6**, 219–229. <https://doi.org/10.1016/j.mex.2019.01.004>
658 Schuman, H., and Presser, S. 1996 *Questions and answers in attitude surveys : experiments on*
659 *question forms, wording, and context*, Academic Press, San Diego, 147–178.
660 Sherry, E., Halseth, R., Fondahl, G., Karjala, M., and Leon, B. 2005 Local-level criteria and
661 indicators: An aboriginal perspective on sustainable forest management, *Forestry*, **78**(5), 513–
662 539. <https://doi.org/10.1093/forestry/cpi048>
663 Shimizu, K., and Saito, H. 2021 Country-wide mapping of harvest areas and post-harvest forest
664 recovery using Landsat time series data in Japan, *Int. J. Appl. Earth Obs. Geoinf.*, **104**, 102555.

665 <https://doi.org/10.1016/J.JAG.2021.102555>

666 Soto, J. R., Escobedo, F. J., Khachatryan, H., and Adams, D. C. 2018 Consumer demand for urban
667 forest ecosystem services and disservices: Examining trade-offs using choice experiments and
668 best-worst scaling, *Ecosyst. Serv.*, **29**, 31–39. <https://doi.org/10.1016/J.ECOSER.2017.11.009>

669 Starbuck, C. M., Berrens, R. P., and McKee, M. 2006 Simulating changes in forest recreation demand
670 and associated economic impacts due to fire and fuels management activities, *For. Policy Econ.*,
671 **8**(1), 52–66. <https://doi.org/10.1016/J.FORPOL.2004.05.004>

672 Stritih, A., Bebi, P., Rossi, C., and Grêt-Regamey, A. 2021 Addressing disturbance risk to mountain
673 forest ecosystem services, *J. Environ. Manage.*, **296**, 113188.
674 <https://doi.org/10.1016/J.JENVMAN.2021.113188>

675 Stumpf, A., and Kerle, N. 2011 Object-oriented mapping of landslides using Random Forests, *Remote*
676 *Sens. Environ.*, **115**(10), 2564–2577. <https://doi.org/10.1016/j.rse.2011.05.013>

677 TEEB 2010 *The Economics of Ecosystems and Biodiversity Ecological and Economic Foundations* (P.
678 Kumar, Ed.), Earthscan, London and Washington, pp. 422.

679 Termansen, M., McClean, C. J., and Jensen, F. S. 2013 Modelling and mapping spatial heterogeneity
680 in forest recreation services, *Ecol. Econ.*, **92**, 48–57.
681 <https://doi.org/10.1016/J.ECOLECON.2013.05.001>

682 Yamada, Y. 2020 Optimization of regional forest planning with multiple decision-makers, *J. For.*
683 *Res.*, **25**(6), 379–388. <https://doi.org/10.1080/13416979.2020.1807694>

684 Yamada, Y., and Yamaura, Y. 2017 Decision Support System for Adaptive Regional-Scale Forest
685 Management by Multiple Decision-Makers, *Forests*, **8**(11), 453.
686 <https://doi.org/10.3390/f8110453>

687 Yamaura, Y., Yamada, Y., Matsuura, T., Tamai, K., Taki, H., Sato, T., Hashimoto, S., Murakami, W.,
688 Toda, K., Saito, H., Nanko, K., Ito, E., Takayama, N., Tsuzuki, N., Takahashi, M., Yamaki, K.,
689 and Sano, M. 2021 Modeling impacts of broad-scale plantation forestry on ecosystem services in
690 the past 60 years and for the future, *Ecosyst. Serv.*, **49**, 101271.
691 <https://doi.org/10.1016/j.ecoser.2021.101271>

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694 **Figures and legends**

695 Fig. 1 Different trade-off relationships at forest stand scales produce different regional trade-off
696 relationships. Pattern A: forest stands with higher non-timber ES value are more likely to be logged;
697 Pattern B: In some cases, the relationships between the use of timber ES and non-timber ES values is
698 unclear; Pattern C: foresters may avoid cutting forest stands with high non-timber ES values.

699 Fig. 2 Study area in Kyushu, southwestern Japan.

700 Fig. 3 Distributions of forest plantations in the study area. Values are proportions of forest plantations
701 in the whole area of the region.

702 Fig. 4 Relative frequency distributions of forest plantation ages for in each category of slope degree
703 (upper row) and distances from roads (lower row). The dashed lines indicate median values.

704 Fig. 5 Distributions of ecosystem service values of (a) landslide susceptibility, (b) sediment retention
705 and (c) forest recreation.

706 Fig. 6 Locations of forests logged in 2015–2019.

707 Fig. 7 Logged area ratios in each category. Each value shows the ratio of logged area to forest
708 plantations in each category (sum = 1.0).

709 Fig. 8 Logged area ratio of each category with (a) landslide susceptibility class (x-axis): a, very stable;
710 b, stable; c, unstable; d, very unstable; (b) annual soil loss class (x-axis): a, slight ($<1 \text{ t ha}^{-1} \text{ year}^{-1}$);
711 b, moderate (1–5); c, high (5–10); d, significant (>10);, and (c) forest recreation index class (x-axis): a,
712 (0); b, (0–0.2); c, (0.2–0.35); d, (>0.35). Letters on the red curves indicate other classes from which
713 each class differed significantly ($P < 0.1$).

714 Fig. 9 Relative frequency distribution of forest plantation ages in each category of landslide
715 susceptibility class (upper row): a, very stable; b, stable; c, unstable; d, very unstable; sediment
716 retention class (middle row): a, slight ($<0.1 \text{ t ha}^{-1} \text{ year}^{-1}$); b, moderate (0.1–1); c, high (1–10); d,
717 significant (>10); or forest recreation class (lower row): a, <0.1 ; b, 0.1–0.25; c, 0.25–0.5; d, ≥ 0.5 . The
718 dashed lines are median values.

719 **Tables**

720 Table 1 Comparison of logging practices and environmental awareness in the three regions.

	Saiki	Bungo-Ono	Taketa
Timber supplier	The forest owners' association	Non-regional forestry companies	Several local forestry companies
Awareness of landslide prevention and sediment retention	High.	Low.	High.
ESs			
Awareness of forest recreation ES	Low.	Low.	Low.
Logging criteria for conservation of ESs	Yes.	No.	No.

721

722 Table 2 Logged area ratio in each region in 2015-2019.

	Saiki	Bungo-Ono	Taketa
	7.24%	3.64%	3.75%

723